

**Working Paper for COMEAP Report “Associations of long-term average concentrations of nitrogen dioxide with mortality” published July 2018**

**Working Paper 1: Systematic review and meta-analysis of cohort studies of NO<sub>2</sub> and all-cause mortality**

**Richard W Atkinson & Barbara K Butland**

# SYSTEMATIC REVIEW AND META-ANALYSIS OF COHORT STUDIES OF NO<sub>2</sub> AND ALL-CAUSE MORTALITY

**RW Atkinson & BK Butland**

## **AIM**

To recommend a coefficient for long-term exposure to NO<sub>2</sub> and all-cause mortality for use in health impact assessment.

To achieve this aim the following objectives were set:

- to undertake a systematic literature search to identify cohort studies reporting hazard ratios (HRs) for long-term exposure to NO<sub>2</sub> and all-cause mortality
- to select studies for meta-analysis and derive a summary HR
- to identify studies reporting HRs for NO<sub>2</sub> and all-cause mortality adjusting for PM<sub>2.5</sub>
- to estimate the reduction in the HR for NO<sub>2</sub> after adjustment for PM<sub>2.5</sub>

## **METHODS**

To identify publications reporting results for cohort studies of NO<sub>2</sub> and mortality we conducted a broad search of the online medical databases EMBASE and MEDLINE, supplemented with citation searches of recently published literature reviews and COMEAP papers.

### **Search strategy**

Three search strings were applied to Ovid Medline (R) without Revisions for the period 1996 to October Week 1 2015 and to Embase for the period 1996 to 2015 Week 41. The search strings were: a) "cohort" & "no2" & "mortality"; b) "cohort" & "air pollution" & "mortality"; and c) "long-term" & "no2" & "mortality". These searches were supplemented by citation searches in 6 review articles.<sup>1-6</sup>

Our search strategy excluded conference abstracts, conference papers, notes, editorials and letters. Cohort studies were selected if they included a 'long-term' exposure metric for NO<sub>2</sub>—studies using daily or monthly exposures were excluded. Cohort studies were also required to have individual-level covariate information. Cross sectional, case control and nested case-control studies were excluded from the review. The outcome studied had to be all-cause or cause-specific mortality (not disease incidence).

### **Inclusion/exclusion criteria**

Studies providing HRs for NO<sub>x</sub> were excluded from the review. Studies providing quantitative HRs (i.e. not graphically) together with either standard errors or 95% confidence intervals were selected. Adequate information had to be provided to allow presentation of such estimates as per 10 µg/m<sup>3</sup> increase in pollutant. Where results for various follow-up periods were provided we chose the follow-up period that was most up to date.

If results for the same outcomes were available for the full cohort or a subset we used results from the full cohort unless these results were considered to be out of date (e.g. statistical analysis, exposure assessment, date of last follow-up). Two studies from the same cohort were only included if they provided results for different outcomes.

### **Data extraction**

Cohort and estimate level information were extracted from each paper/online supplement. These data included cohort name, country, cohort description, date of enrolment of cohort members, age at enrolment, number of subjects, follow-up period, exposure period and exposure assessment method (measured/modelled).

All HRs were standardised to 10µg/m<sup>3</sup> increase in NO<sub>2</sub>. Where the units used in the original study were ppb, a conversion factor of 1.88µg/m<sup>3</sup> per 1ppb was used (assuming 25°C and 1013mb atmospheric pressure).

### **Quality criteria**

Covariate adjustment should include individual-level age, sex, smoking and BMI. Also adjustment for some marker of socioeconomic status (e.g. education level, income etc.) at either the individual or ecological level.

### **Meta-analysis**

All analyses were conducted in STATA Version 12. All studies reported HRs together with 95% confidence intervals. Therefore, estimates of the standard error were derived using each limit value in turn and the two estimates averaged. Forest plots were used to display study information and HRs graphically. Meta-analytic summary estimates were calculated using fixed/random effects models using the program 'metan' in STATA. Heterogeneity was assessed using the I<sup>2</sup> statistic. Small study bias was assessed using Begg<sup>7</sup> and Egger<sup>8</sup> tests and the Trim and Fill procedure<sup>9</sup>.

## RESULTS

### Literature search

996 records were identified from the database searches and other sources. After removal of duplicates and application of the inclusion/exclusion criteria, 66 articles were identified for full-text review. After exclusion of studies that reported results for NO<sub>x</sub> (n=4), replicated results reported elsewhere (n=6), or did not report HRs quantitatively (n=11), 45 articles remained, of which 20 did not adjust fully for all confounders.

Causes of 'all' deaths were variously described as 'All cause', 'Natural causes' and 'Non Accidental' and were re-coded as 'All Cause'. Studies only of cause-specific deaths were not considered further. 28 articles analysing 21 cohorts (including the ESCAPE consortium of individual cohorts) reported results for all-cause mortality.

Cohorts comprising selected subgroups defined by pre-existing disease were excluded as being unrepresentative of the general population and therefore unsuitable for the purpose of a health impact assessment exercise. Consequently, 6 publications from 5 cohorts (stroke survivors<sup>10</sup>; CHD survivors<sup>11</sup>; attendees at respiratory clinic<sup>12</sup>; ACS survivors<sup>13</sup>; and hypertensive US veterans<sup>14,15</sup>) were excluded.

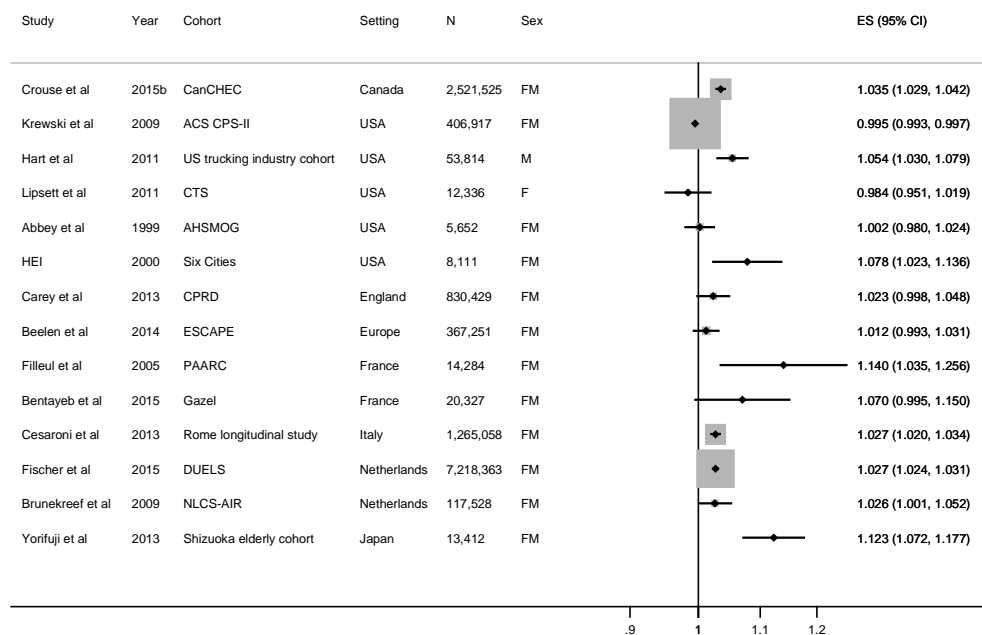
Three studies<sup>16-18</sup> were excluded as their results were included in ESCAPE meta-analysis<sup>19</sup>. The ESCAPE study provided a meta-analytical result (of cohorts in the ESCAPE project) only and not individual cohort HRs.

A further 5 studies were excluded as the same cohorts were analysed in other publications included in the review.<sup>20-24</sup> In one study<sup>25</sup>, results for two cohorts were reported – the HR for the ACS CPS II cohort reported in this study was not used.

### Meta-analysis

Following these exclusions, results from 14 separate cohorts (including the ESCAPE consortium of 22 individual cohorts) reported results for NO<sub>2</sub> and all-cause mortality.<sup>19,25-37</sup>

The majority of the 14 cohorts were in European populations (7 including the ESCAPE study); 6 cohorts were from North America and a single cohort from Japan. Key cohort characteristics and corresponding HRs are presented in Figure 1. There was substantial heterogeneity between effect estimates, I<sup>2</sup>=96%.



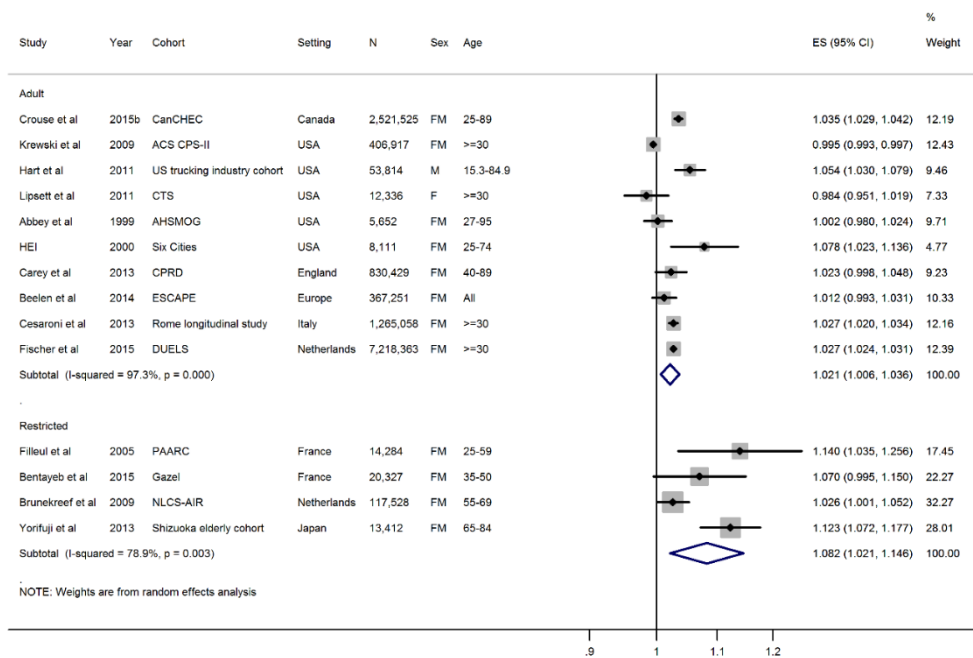
**Figure 1 HRs (95% CI) per 10µg/m<sup>3</sup> for cohort studies reporting associations between NO<sub>2</sub> and all-cause mortality**

Analysis stratified by adults across a broad age range vs specific age groups is shown in Figure 2. Four studies focused on specific age groups. The summary HR for these studies was substantially larger than for studies with broader age ranges at recruitment. As the focus of our review was to derive a summary HR considered to be representative of the general population we excluded, from further analyses, the cohort studies (n=4) in adults restricted to narrow age ranges at cohort entry.<sup>1</sup>

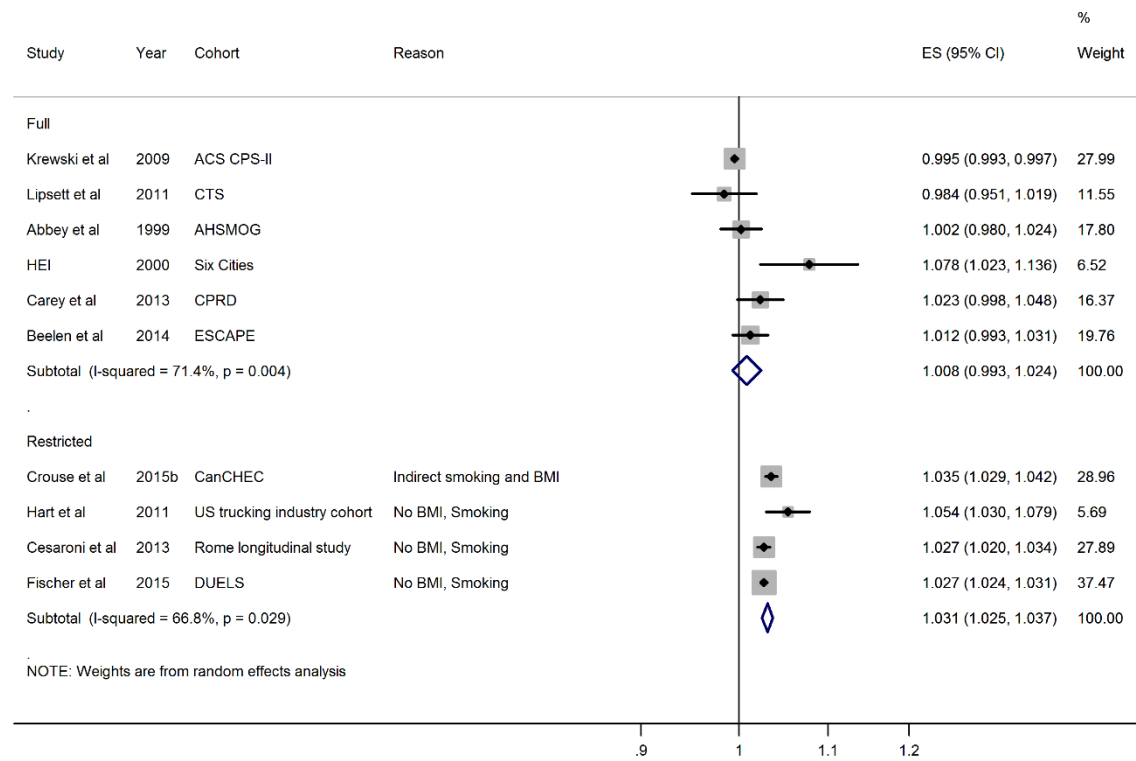
When the remaining cohorts (n=10) were stratified by level of covariate adjustment – i.e. those controlling for the required confounding factors and those that did not (Figure 3) there was a substantial difference in the NO<sub>2</sub> random-effects summary estimates; 1.008 (95% CI: 0.993, 1.024) vs 1.031 (95% CI: 1.025, 1.037) per 10µg/m<sup>3</sup> respectively.<sup>2</sup>

<sup>1</sup> Bentayeb et al 2015 does not report any age restriction on cohort members at recruitment. In the original meta-analysis conducted in 2015 this study was therefore coded as 'adult' rather than 'restricted'. The meta-analytical summary estimate was then used in subsequent health impact calculations. On further investigation conducted in July 2017, a related paper was identified which indicated the age range of cohort participants in Bentayeb et al 2015 was restricted to ages 35-50. The coding for Bentayeb et al 2015 was changed therefore to reflect this new information, and the meta-analyses presented in this working paper use this coding.

<sup>2</sup> The recoding of Bentayeb et al 2015 did not alter materially this finding



**Figure 2 HRs (95% CI) for cohort studies reporting associations between NO<sub>2</sub> and all-cause mortality stratified by age groups**

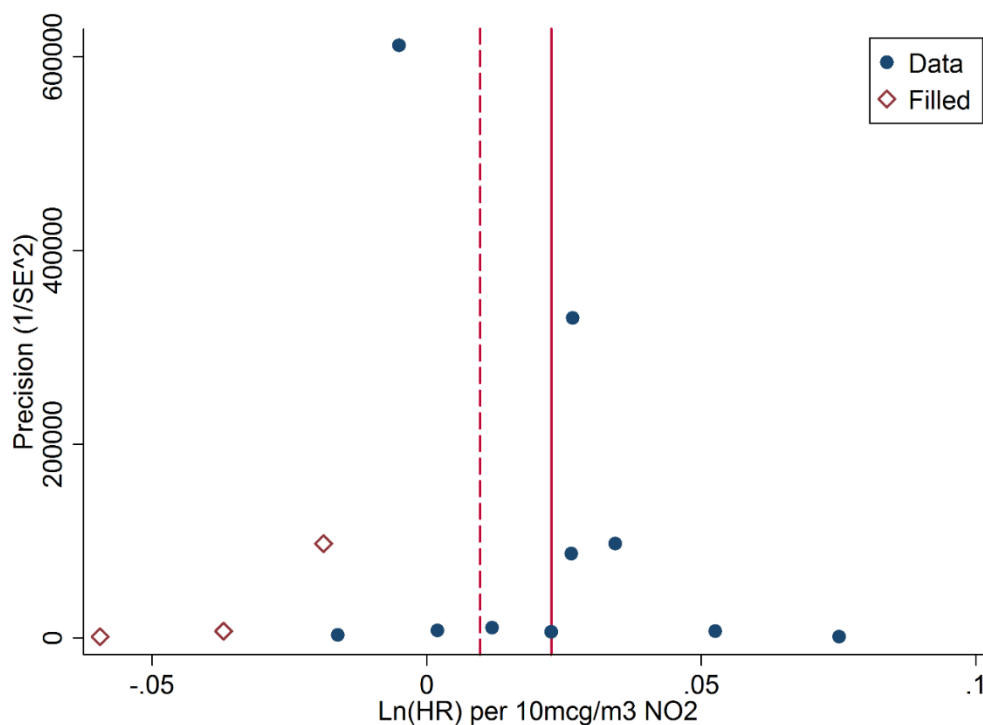


**Figure 3 HRs (95% CI) per 10µg/m<sup>3</sup> for cohort studies reporting associations between NO<sub>2</sub> and all-cause mortality stratified by covariate adjustment**

Based upon the 10 studies presented in Figure 3, the fixed-effects summary HR for NO<sub>2</sub> was 1.010, (95% CI: 1.009, 1.012) per 10µg/m<sup>3</sup>. There was substantial heterogeneity between estimates, I<sup>2</sup>=97%. The corresponding random-effects summary HR was 1.021 (95% CI: 1.006, 1.036) per 10µg/m<sup>3</sup>.<sup>3</sup>

### Small study bias

P-values for Begg's and Egger's tests were 0.37 and 0.32 respectively. Application of the trim and fill technique indicated the need to impute 3 additional study estimates (Figure 4) to adjust for small study bias assuming a fixed-random effects model, adjusted HR=1.010 (95% CI: 0.996, 1.023). Assuming a random-random effects model the Trim & Fill procedure did not indicate the need to impute additional estimates to achieve symmetry.



**Figure 4 Funnel plot showing study precision against Ln(HR) using fixed-random effects and summary estimates (Solid line = unadjusted HR, Dash line = adjusted HR)**

<sup>3</sup> The incorrect coding of age range for Bentayeb et al. 2015 at the start of our review meant that the original meta-analysis was conducted on 11 studies. This analysis reported a fixed-effects summary HR of 1.010, (95% CI: 1.009, 1.012) per 10 µg/m<sup>3</sup> with substantial heterogeneity between estimates, I<sup>2</sup>=97%. The corresponding random-effects summary HR was 1.023 (95% CI: 1.008, 1.037) per 10 µg/m<sup>3</sup>.

### **Independence from PM<sub>2.5</sub>**

Table 1 and Figure 5 show the HRs (95% CI) from single- and two-pollutant models for NO<sub>2</sub> and PM in the cohorts reporting results for all-cause mortality. Five studies adjusted for PM<sub>2.5</sub> and one for PM<sub>10</sub>. Confidence intervals for unadjusted and adjusted HRs overlapped. None of the cohorts assessed the independence of NO<sub>2</sub> from traffic related particles (ultrafine, elemental carbon etc.).

The percentage reduction in the ln(HR) for NO<sub>2</sub> after adjustment for PM varied from 10% to 95% and in one study, a negative association between NO<sub>2</sub> and mortality reduced further upon adjustment for PM<sub>2.5</sub>. For PM (PM<sub>2.5</sub> or PM<sub>10</sub>) the reductions were between 0% and 82% and in one study the PM<sub>2.5</sub> coefficient increased by 42% upon adjustment for NO<sub>2</sub>. Correlations between NO<sub>2</sub> and PM<sub>2.5</sub> (Table 1) were high in 2 studies (0.79 and 0.85), moderate in 2 studies (0.2-0.7 and 0.55) and weak (-0.08) in one study. In the single study using PM<sub>10</sub>, rather than PM<sub>2.5</sub>, the correlation with NO<sub>2</sub> was 0.58.



Table 1 Hazard ratios (HR) from single and two pollutant models for NO<sub>2</sub> and PM<sub>2.5</sub> or PM<sub>10</sub> (HRs are expressed per IQR)

Study	Cohort	Corr NO <sub>2</sub> /PM <sub>2.5</sub>	NO <sub>2</sub> IQR (µg/m <sup>3</sup> )	HR NO <sub>2</sub>	NO <sub>2</sub> adj PM <sub>2.5</sub> /PM <sub>10</sub>	% <sup>5</sup>	PM <sub>2.5</sub> /PM <sub>10</sub> IQR (µg/m <sup>3</sup> )	HR PM <sub>2.5</sub> /PM <sub>10</sub>	PM <sub>2.5</sub> /PM <sub>10</sub> adj NO <sub>2</sub>	% <sup>5</sup>	Combined NO <sub>2</sub> adj/PM adj HR
Cesaroni et al 2013	Rome	0.79	10.7	1.029 (1.022, 1.036)	1.026 (1.015, 1.037)	<b>10</b>	5.7	1.023 (1.016, 1.031)	1.004 (0.994, 1.015)	<b>82</b>	1.030
Carey et al 2013 <sup>1</sup>	CPRD	0.85	10.7	1.022 (0.995, 1.049)	1.001 (0.959, 1.044)	<b>95</b>	1.9	1.023 (1.000, 1.046)	1.023 (0.989, 1.060)	<b>0</b>	1.024
Beelen et al 2014 <sup>2</sup>	ESCAPE	0.2-<0.7	10.0	1.015 (0.993, 1.036)	1.007 (0.967, 1.049)	<b>53</b>	5.0	1.070 (1.016, 1.127)	1.060 (0.977, 1.150)	<b>14</b>	1.067
Fischer et al 2015 <sup>3</sup>	DUELS	0.58	10.0	1.027 (1.023, 1.030)	1.019 (1.015, 1.023)	<b>29</b>	2.4	1.019 (1.016, 1.022)	1.010 (1.007, 1.013)	<b>46</b>	1.029
HEI 2000 <sup>4</sup>	ACS CPS II	-0.08	81.4	0.95 (0.89, 1.01)	0.90 (0.84, 0.96)	<b>105</b>	24.5	1.15 (1.05, 1.25)	1.22 (1.11, 1.33)	<b>-42</b>	1.09
Jerret et al 2013	ACS CPS II	0.55	7.7	1.031 (1.008, 1.056)	1.025 (0.997, 1.054)	<b>19</b>	5.3	1.032 (1.002, 1.062)	1.015 (0.980, 1.050)	<b>53</b>	1.040

Notes:

1 PM<sub>2.5</sub> results –personal communication.

2 Based on 14 cohorts in which correlation between NO<sub>2</sub> and PM<sub>2.5</sub> was less than 0.7. HRs are presented per 10 µg/m<sup>3</sup> NO<sub>2</sub> and 5 µg/m<sup>3</sup> PM<sub>2.5</sub>

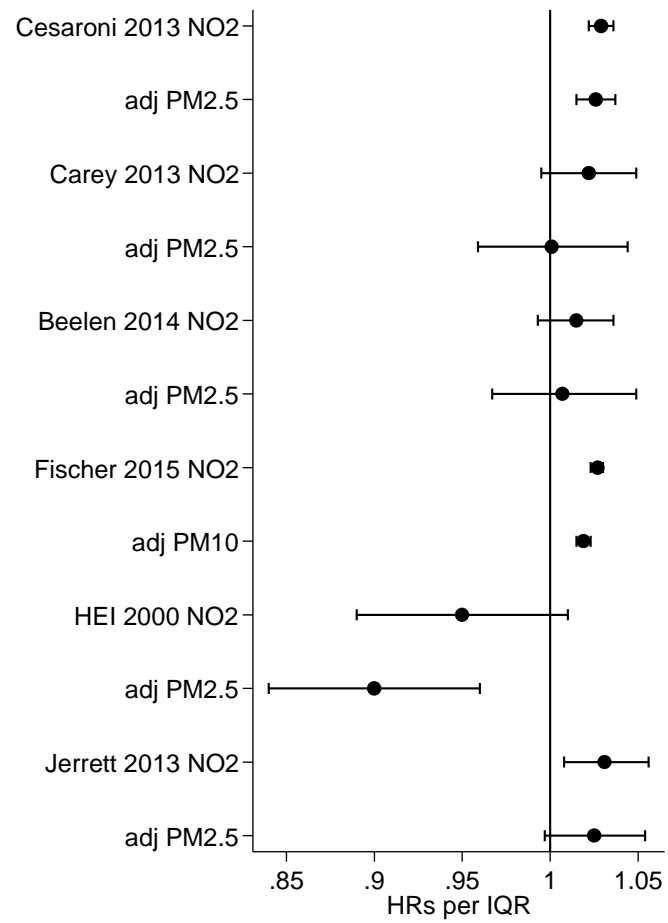
3 PM<sub>10</sub>

4 HR (95% CI) for min-max range of average concentrations in fine particulate cohort (41 cities).

5 % reduction in ln(HR)

(HR reported to 3 decimal places taken from publication or provided by personal communication)

A



B

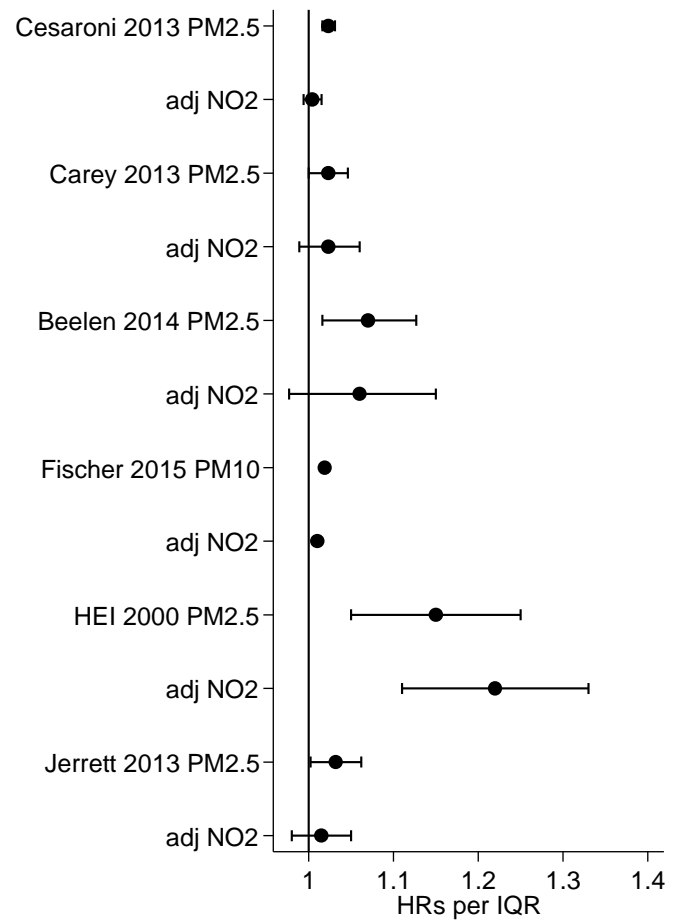


Figure 5 Hazard ratios (95% CI) from single- and two-pollutant models for NO<sub>2</sub> (A) and PM<sub>2.5</sub> or PM<sub>10</sub> (B) (HRs are expressed per IQR or selected increments in multi-centre studies)

## DISCUSSION

### 1. Adjustment for individual confounders

Having excluded four cohorts with a restricted age range, four of the remaining 10 studies selected for meta-analysis did not control for individual measures of smoking and BMI. When stratified by level of covariate adjustment – i.e. those controlling for the required confounding factors and those that did not (Figure 3) the HRs differed substantially: 1.008 vs 1.031 per 10  $\mu\text{g}/\text{m}^3$  respectively. Both Cesaroni et al 2013 and Fischer et al 2015 note this limitation of their studies. Cesaroni et al presented results from a small subset (7845) for which individual smoking measures were available and noted that adjustment for smoking did not alter associations between  $\text{NO}_2$  and mortality. Cesaroni et al also adjusted for smoking related comorbidities. Fischer et al conducted a sensitivity analysis adjusted for regional age-standardised smoking-attributable mortality and noted an attenuation of the association from 1.03 (95% CI: 1.02, 1.03) to 1.02 (95% CI: 1.02, 1.03). A sensitivity analysis using the English cohort found that adjustment for individual level smoking status and BMI after adjustment for a small area marker of socio economic status attenuated the HRs by a further 15% (personal communication). The possibility remains, therefore, that studies unable to control for individual confounders may be overstating the size of the association between long-term  $\text{NO}_2$  and all-cause mortality.

### 2. Heterogeneity

For the 10 single pollutant HRs selected for meta-analysis, the fixed- and random-effects summary estimates differed substantially: 1.010 (95% CI: 1.009, 1.012) and 1.021 (95% CI: 1.006, 1.036) per 10  $\mu\text{g}/\text{m}^3$  respectively. Under the fixed-effects model, all studies are assumed to estimate a common HR. In the meta-analysis therefore, only study precision determines study weight. In a random-effects model however, it is assumed that the study populations (and the study methodology) can differ in ways that can impact on the estimated HRs and a distribution of HRs is assumed. Weights in a random-effects meta-analysis are determined not just by study size, but also between-study variance. Hence, greater weight can be given to smaller studies and less weight to larger studies. In this meta-analysis, 97% of the variation in the HRs was attributable to between-study variance. The two models have different conceptual frameworks and when interpreting their results, it is important to understand possible reasons for the between-study variability. These may relate to population characteristics, baseline population risk, exposure assessment concentrations and sources of co-pollutants and variability in model specification including potential confounders. Careful interpretation of the model results are warranted therefore.

### **3. Small study bias**

Small study bias encompasses publication bias – the publication of adverse, imprecise study results. Publication bias can arise from a number of stages in the process of publication of research findings. These include analyst decisions in model selection and the reporting of null results, decisions by study investigators to submit results for peer review and decisions by journal editors to publish study findings. Small study bias can also be due to heterogeneity between studies and differences in study methodology. It can be identified using a number of graphical and statistical tests.<sup>7-9</sup> The presence of small study bias in air pollution epidemiology has been noted previously.<sup>38</sup>

The results from the Begg and Egger tests reported non-significant results whilst the Trim and Fill procedure required the imputation of additional results to achieve symmetry suggesting the presence of small study bias. The degree of adjustment to the summary estimate varied under the different model specifications available within the Trim and Fill technique. The performance of the Trim and Fill procedure, especially in the presence of heterogeneity between study estimates, has been assessed in simulation studies<sup>39,40</sup> Given the substantial heterogeneity between our study estimates the interpretation of the results from the procedure require further investigation. We therefore do not recommend adjustment for small study bias until further assessment of the causes of heterogeneity have been identified but note the possibility that the unadjusted HR may be subject to some bias as a result.

### **4. Multi-pollutant models**

The difficulty in interpreting regression coefficients for correlated variables in multivariate regression is well documented in the statistical literature. More recently, the difficulties in interpreting coefficients in multi-pollutant models has received attention.<sup>41,42</sup> These difficulties include: 1) correlation between pollutants (arising due to common sources and meteorological conditions) can lead to unstable parameter estimation; 2) differential measurement error between pollutants can lead to the ‘transfer’ of an association from the less well measured (but true) pollutant to the better measured (but incorrect) pollutant; and 3) statistical models do not generally assess interactions between pollutants and these assessments are required to interpret correctly model main effects.

Given the characteristics of the 6 studies reporting multi-pollutant model results and the problems in interpreting coefficients from multi-pollutant models, the validity of the adjusted coefficients is questionable. Table 1 also shows the combined HRs for the two adjusted pollutant coefficients (i.e. NO<sub>2</sub> adjusted for PM and PM adjusted for NO<sub>2</sub>). In four studies the

combined HRs were similar to the NO<sub>2</sub> or PM single pollutant HRs. In one study the combined NO<sub>2</sub> and PM HRs was larger than both single pollutant model HRs and in another the combined HR lay between the two single pollutant HRs. The combined HRs provide more stable and reliable estimates of the associations between exposure to the two correlated pollutants and all-cause mortality (notwithstanding the lack of interaction terms). These combined estimates could be used in formulating a multi-pollutant approach to regulatory policy as advocated by Greenbaum and colleagues.<sup>43</sup>

## REFERENCES

1. Hoek G, Krishnan RM, Beelen R, et al. Long-term air pollution exposure and cardio-respiratory mortality: a review. *Environmental health : a global access science source* 2013; **12**(1): 43.
2. Latza U, Gerdes S, Baur X. Effects of nitrogen dioxide on human health: Systematic review of experimental and epidemiological studies conducted between 2002 and 2006. *International Journal of Hygiene and Environmental Health* 2009; **212**(3): 271-87.
3. Brunekreef B. Health effects of air pollution observed in cohort studies in Europe. *Journal of Exposure Science and Environmental Epidemiology* 2007; **17**(SUPPL. 2): S61-S5.
4. Faustini A, Rapp R, Forastiere F. Nitrogen dioxide and mortality: Review and meta-analysis of long-term studies. *European Respiratory Journal* 2014; **44**(3): 744-53.
5. Hamra GB, Laden F, Cohen AJ, Raaschou-Nielsen O, Brauer M, Loomis D. Lung Cancer and Exposure to Nitrogen Dioxide and Traffic: A Systematic Review and Meta-Analysis. *Environ Health Perspect* 2015; **123**(11): 1107-12.
6. Atkinson RW, Butland BK, Dimitroulopoulou C, et al. Long-term exposure to ambient ozone and mortality: a quantitative systematic review and meta-analysis of evidence from cohort studies. *BMJ Open* 2016; **6**(2): e009493.
7. Begg CB, Mazumdar M. Operating characteristics of a rank correlation test for publication bias. *Biometrics* 1994; **50**(4): 1088-101.
8. Egger M, Davey Smith G, Schneider M, Minder C. Bias in meta-analysis detected by a simple, graphical test. *BMJ* 1997; **315**(7109): 629-34.
9. Duval S, Tweedie R. Trim and fill: A simple funnel-plot-based method of testing and adjusting for publication bias in meta-analysis. *Biometrics* 2000; **56**(2): 455-63.
10. Maheswaran R, Pearson T, Smeeton NC, Beevers SD, Campbell MJ, Wolfe CD. Impact of outdoor air pollution on survival after stroke: Population-based cohort study. *Stroke* 2010; **41**(5): 869-77.
11. Rosenlund M, Picciotto S, Forastiere F, Stafoggia M, Perucci CA. Traffic-related air pollution in relation to incidence and prognosis of coronary heart disease. *Epidemiology* 2008; **19**(1): 121-8.
12. Jerrett M, Finkelstein MM, Brook JR, et al. A cohort study of traffic-related air pollution and mortality in Toronto, Ontario, Canada. *Environmental Health Perspectives* 2009; **117**(5): 772-7.
13. Tonne C, Wilkinson P. Long-term exposure to air pollution is associated with survival following acute coronary syndrome. *European Heart Journal* 2013; **34**(17): 1306-11.
14. Lipfert F, Baty J, Miller J, Wyzga R. PM2.5 constituents and related air quality variables as predictors of survival in a cohort of U.S. military veterans. *Inhalation Toxicology* 2006; **18**(9): 645-57.
15. Lipfert FW, Wyzga RE, Baty JD, Miller JP. Traffic density as a surrogate measure of environmental exposures in studies of air pollution health effects: Long-term mortality in a cohort of US veterans. *Atmospheric Environment* 2006; **40**(1): 154-69.
16. Raaschou-Nielsen O, Andersen ZJ, Jensen SS, et al. Traffic air pollution and mortality from cardiovascular disease and all causes: A Danish cohort study. *Environmental Health: A Global Access Science Source* 2012; **11**(1).
17. Gehring U, Heinrich J, Kramer U, et al. Long-term exposure to ambient air pollution and cardiopulmonary mortality in women. *Epidemiology* 2006; **17**(5): 545-51.
18. Heinrich J, Thiering E, Rzehak P, et al. Long-term exposure to NO2 and PM10 and all-cause and cause-specific mortality in a prospective cohort of women. *Occupational and Environmental Medicine* 2013; **70**(3): 179-86.
19. Beelen R, Raaschou-Nielsen O, Stafoggia M, et al. Effects of long-term exposure to air pollution on natural-cause mortality: An analysis of 22 European cohorts within the multicentre ESCAPE project. *The Lancet* 2014; **383**(9919): 785-95.

20. Yorifuji T, Kashima S, Tsuda T, et al. Long-term exposure to traffic-related air pollution and mortality in Shizuoka, Japan. *Occupational and Environmental Medicine* 2010; **67**(2): 111-7.
21. Cesaroni G, Porta D, Badaloni C, et al. Nitrogen dioxide levels estimated from land use regression models several years apart and association with mortality in a large cohort study. *Environmental Health: A Global Access Science Source* 2012; **11**(1).
22. Hoek G, Brunekreef B, Goldbohm S, Fischer P, Van Den Brandt PA. Association between mortality and indicators of traffic-related air pollution in the Netherlands: A cohort study. *Lancet* 2002; **360**(9341): 1203-9.
23. Jerrett M, Burnett RT, Beckerman BS, et al. Spatial analysis of air pollution and mortality in California. *American Journal of Respiratory and Critical Care Medicine* 2013; **188**(5): 593-9.
24. Crouse DL, Peters PA, Villeneuve PJ, et al. Within- and between-city contrasts in nitrogen dioxide and mortality in 10 Canadian cities; a subset of the Canadian Census Health and Environment Cohort (CanCHEC). *J Expo Sci Environ Epidemiol* 2015; **25**(5): 482-9.
25. HEI. Reanalysis of the Harvard Six Cities Study and the American Cancer Society Study of Particulate Air Pollution and Mortality: A Special Report of the Institute's Particle Epidemiology Reanalysis Project., 2000.
26. Crouse DL, Peters PA, Hystad P, et al. Ambient PM<sub>2.5</sub>, O<sub>3</sub>, and NO<sub>2</sub> Exposures and Associations with Mortality over 16 Years of Follow-Up in the Canadian Census Health and Environment Cohort (CanCHEC). *Environ Health Perspect* 2015; **123**(11): 1180-6.
27. Krewski D, Jerrett M, Burnett RT, et al. Extended follow-up and spatial analysis of the American Cancer Society study linking particulate air pollution and mortality. Research report (Health Effects Institute) 2009; (140): 5-114; discussion 5-36.
28. Abbey DE, Nishino N, McDonnell WF, et al. Long-term inhalable particles and other air pollutants related to mortality in nonsmokers. *Am J Respir Crit Care Med* 1999; **159**(2): 373-82.
29. Lipsett MJ, Ostro BD, Reynolds P, et al. Long-term exposure to air pollution and cardiorespiratory disease in the California teachers study cohort. *American Journal of Respiratory and Critical Care Medicine* 2011; **184**(7): 828-35.
30. Brunekreef B, Beelen R, Hoek G, Schouten L, Bausch-Goldbohm S, Fischer P et al. Effects of long-term exposure to traffic-related air pollution on respiratory and cardiovascular mortality in the Netherlands: the NLCS-AIR study. Research report (Health Effects Institute). 2009(139)
31. Hart JE, Garshick E, Dockery DW, Smith TJ, Ryan L, Laden F. Long-term ambient multipollutant exposures and mortality. *American Journal of Respiratory and Critical Care Medicine* 2011; **183**(1): 73-8.
32. Carey IM, Atkinson RW, Kent AJ, van ST, Cook DG, Anderson HR. Mortality associations with long-term exposure to outdoor air pollution in a national English cohort. *AmJRespirCrit Care Med* 2013; **187**(11): 1226-33.
33. Filleul L, Rondeau V, Vandentorren S, et al. Twenty five year mortality and air pollution: Results from the French PAARC survey. *Occupational and Environmental Medicine* 2005; **62**(7): 453-60.
34. Cesaroni G, Badaloni C, Gariazzo C, et al. Long-term exposure to urban air pollution and mortality in a cohort of more than a million adults in Rome. *Environmental Health Perspectives* 2013; **121**(3): 324-31.
35. Fischer PH, Marra M, Ameling CB, et al. Air pollution and mortality in seven million adults: The dutch environmental longitudinal study (DUELS). *Environmental Health Perspectives* 2015; **123**(7): 697-704.
36. Yorifuji T, Kashima S, Tsuda T, et al. Long-term exposure to traffic-related air pollution and the risk of death from hemorrhagic stroke and lung cancer in Shizuoka, Japan. *Science of the Total Environment* 2013; **443**: 397-402.

37. Bentayeb M, Wagner V, Stempfelet M, et al. Association between long-term exposure to air pollution and mortality in France: A 25-year follow-up study. *Environment International* 2015; **85**: 5-14.
38. Anderson HR, Atkinson RW, Peacock JL, Sweeting MJ, Marston L. Ambient particulate matter and health effects - Publication bias in studies of short-term associations. *Epidemiology* 2005; **16**(2): 155-63.
39. Peters JL, Sutton AJ, Jones DR, Abrams KR, Rushton L. Performance of the trim and fill method in the presence of publication bias and between-study heterogeneity. *Stat Med* 2007; **26**(25): 4544-62.
40. Terrin N, Schmid CH, Lau J, Olkin I. Adjusting for publication bias in the presence of heterogeneity. *Stat Med* 2003; **22**(13): 2113-26.
41. Dominici F, Peng RD, Barr CD, Bell ML. Protecting human health from air pollution: shifting from a single-pollutant to a multipollutant approach. *Epidemiology* 2010; **21**(2): 187-94.
42. Dionisio KL, Baxter LK, Chang HH. An empirical assessment of exposure measurement error and effect attenuation in bipollutant epidemiologic models. *Environ Health Perspect* 2014; **122**(11): 1216-24.
43. Greenbaum D, Shaikh R. First steps toward multipollutant science for air quality decisions. *Epidemiology* 2010; **21**(2): 195-7.